

Review of methods addressing freshwater use in life cycle inventory and impact assessment

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Abstract

Purpose In recent years, several methods have been developed which propose different freshwater use inventory schemes and impact assessment characterization models considering various cause–effect chain relationships. This work reviewed a multitude of methods and indicators for

freshwater use potentially applicable in life cycle assessment (LCA). This review is used as a basis to identify the key elements to build a scientific consensus for operational characterization methods for LCA.

Methods This evaluation builds on the criteria and procedure developed within the International Reference Life

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Cycle Data System Handbook and has been adapted for the purpose of this project. It therefore includes (1) description of relevant cause–effect chains, (2) definition of criteria to evaluate the existing methods, (3) development of sub-criteria specific to freshwater use, and (4) description and review of existing methods addressing freshwater in LCA. *Results and discussion* No single method is available which comprehensively describes all potential impacts derived from freshwater use. However, this review highlights several key findings to design a characterization method encompassing all the impact pathways of the assessment of freshwater use and consumption in life cycle assessment framework as the following: (1) in most of databases and methods, consistent freshwater balances are not reported either because output is not considered or because polluted freshwater is recalculated based on a critical dilution approach; (2) at the midpoint level, most methods are related to water scarcity index and correspond to the methodological choice of an indicator simplified in terms of the number of parameters (scarcity) and freshwater uses (freshwater consumption or freshwater withdrawal) considered. More comprehensive scarcity indices distinguish different freshwater types and functionalities. (3) At the endpoint level, several methods already exist which report results in units compatible with traditional human health and ecosystem quality damage and cover various cause–effect chains, e.g., the decrease of terrestrial biodiversity due to freshwater consumption. (4) Midpoint and endpoint indicators have various levels of spatial differentiation, i.e., generic factors with no differentiation at all, or country, watershed, and grid cell differentiation.

Conclusions Existing databases should be (1) completed with input and output freshwater flow differentiated according to

water types based on its origin (surface water, groundwater, and precipitation water stored as soil moisture), (2) regionalized, and (3) if possible, characterized with a set of quality parameters. The assessment of impacts related to freshwater use is possible by assembling methods in a comprehensive methodology to characterize each use adequately.

Keywords Ecosystem quality · Freshwater use · Human health · Life cycle assessment · Method review · Resources

1 Introduction

Water is a vital natural resource for all ecosystems, human well-being, and many economic activities. Because of the combination of population growth and economic development leading to increasing human freshwater use (Vörösmarty et al. 2000) and enhanced climate change effects on the global water cycle, water scarcity is becoming an increasing environmental concern. Although freshwater is a local resource, water scarcity is leading to the threat of a global water crisis, with a large share of global population being affected (World Water Assessment Programme UN 2009). Given the actual estimates of global freshwater consumption around 2,600 km³/year and a proposed planetary boundary of 4,000 km³/year consumptive surface and groundwater use (Rockström et al. 2009), it appears that the humanity's freshwater use is currently within the safe operating limit (Shen et al. 2008; Alcamo et al. 2007). Other sources estimate the actual water withdrawal as less than 10 % of the maximum available renewable freshwater resource (Oki and Kanae 2006). However, when considering the regional nature of freshwater scarcity, the majority of global freshwater withdrawals currently takes place in watersheds already experiencing high water scarcity (Ridoutt and Pfister 2010a). According to Ridoutt and Pfister (2010a), the humanity's water footprint (referred as the sum of withdrawals multiplied by local water stress indices) must be globally reduced by approximately 50 % to achieve a sustainable water use. The strong bond between water use and other global environmental and societal systems at various spatial scales such as land use, climate change, and demographic developments justifies both global and regional perspectives for water management to tackle water-related problems (Hoff 2009; Hoekstra 2011).

To tackle this major environmental concern, various initiatives were recently launched in order to develop and standardize analytical tools to measure and assess freshwater use at regional and global scale and to improve the overall management of freshwater resources as well as the overall environmental performance of products and operations. Among these initiatives are the Water Footprint Network (WFN) (Hoekstra et al. 2011), the International Organization for Standardization (ISO) water footprint working group (ISO/TC207/SC5/WG8, ISO 14046 draft), and the World Business Council for

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Sustainable Development (WBCSD 2010) who launched the Global Water Tool and the United Nations Environment Programme (UNEP)/Society of Environmental Toxicology and Chemistry (SETAC) Life Cycle Initiative's working group on the assessment of freshwater use and consumption in life cycle assessment (LCA) called "WULCA" (Koehler and Aoustin 2008). These initiatives also aim to set up public–private partnerships to assist companies in the implementation of water sustainability policies (CEO Water Mandate 2007) and to develop certification programs (AWS 2009).

The authors of this article are part of the UNEP/SETAC Life Cycle Initiative's WULCA working group (Koehler and Aoustin 2008), which involves academic and industrial partners from around the globe who cooperate on the development of methodologies related to freshwater use from a life cycle perspective, including both appropriate freshwater accounting schemes and impact assessment methods. Guidance is provided as scientific consensus regarding the consideration of freshwater in life cycle inventory (LCI) and the choice of life cycle impact assessment (LCIA) methods. Quantitative comparison of LCIA methods will be provided as next step of this work. The working group's deliverables are also used as methodological input to the ISO 14046 water footprint standardization process. A prominent achievement of WULCA was a general framework for the consideration of freshwater resources within LCA (Bayart et al. 2010). This work provided recommendations on freshwater use modeling and relevant impact categories building on the achievements of phase 1 of the UNEP/SETAC Life Cycle Initiative (Bauer et al. 2007) and the conceptual framework including first indicators introduced by Owens (2001).

In the past, most LCA studies did not consider freshwater use, and LCI databases reported freshwater use inventory by determining the total freshwater input from nature or respective technical systems (e.g., drinking water networks) while generally neglecting the water outputs from the LCA system under study (Koehler 2008). LCIA methods applied the amount of freshwater used without characterization factor to address the impacts. Recently, new methodologies were developed which propose freshwater use inventory schemes (Boulay et al. 2011a; Peters et al. 2010; WBCSD 2010) and assess the potential environmental impacts of freshwater use considering various cause–effect relationships (Milà i Canals et al. 2009; Motoshita et al. 2010b; Pfister et al. 2009).

A selection of scientific methods for freshwater use assessment in LCA was evaluated by Berger and Finkbeiner (2010) regarding the methods' scope, input data requirements, and the ISO compliance summarizing the methodological differences. Considering the latter study, the WULCA working group has performed an extensive analysis of a broader variety of freshwater use assessment schemes and metrics applied both in the field of life cycle assessment and water management. In contrast to Berger and Finkbeiner (2010), this work

employs a detailed and systematic analysis to understand differences and similarities in modeling choices using a comprehensive set of evaluation criteria including scientific robustness, transparency and reproducibility, applicability, the level of documentation, and stakeholder acceptance. It is based on the International Reference Life Cycle Data System (ILCD) (JRC-IES 2011).

The goal of the current method review is to provide (1) a comprehensive overview of existing and applicable inventory and impact assessment methods that address freshwater use in a life cycle perspective, (2) an analysis of each method with a set of predefined criteria in order to highlight and understand similarities and differences, (3) an analysis of key parameters to be considered in a consensus-based operational characterization method encompassing the WULCA framework (Bayart et al. 2010), and (4) preliminary application recommendations for practitioners given current state of the art. This study comprises methods for inventorying the use of different freshwater resources as well as for assessing the associated impacts. Methods assessing specific impact of pollutants, i.e., aquatic ecotoxicity, human toxicity, aquatic eutrophication, and aquatic acidification, as well as the recent method dealing with impacts of changed freshwater temperatures due to cooling freshwater discharges (Verones et al. 2010) are not included in this work as they are generally assessed in conventional impact categories of LCA or oriented towards quality-related impact.

2 Methods

The review scheme adopted relies on the approach taken by the European Commission within the International Reference Life Cycle Data System defining the "framework and requirements for LCIA models and indicators" (JRC-IES 2011). The following procedure was followed for the methods review: (1) description of relevant cause–effect chains, (2) definition of criteria to evaluate the existing methods, (3) development of sub-criteria specific to freshwater use, and (4) description and review of existing freshwater use assessment methods.

2.1 Description of relevant cause–effect chains

Figure 1 depicts the cause–effect chains that link freshwater type and use to potential impacts at the mid- and endpoint level and, ultimately, to the related area of protection of human health, ecosystem quality, and resources (Jolliet et al. 2004). The identified cause–effect chains serve as basis for the development of specific criteria linked to freshwater use. In nature, precipitation water (liquid or solid) is differentiated in three types of water that are interconnected: (1)

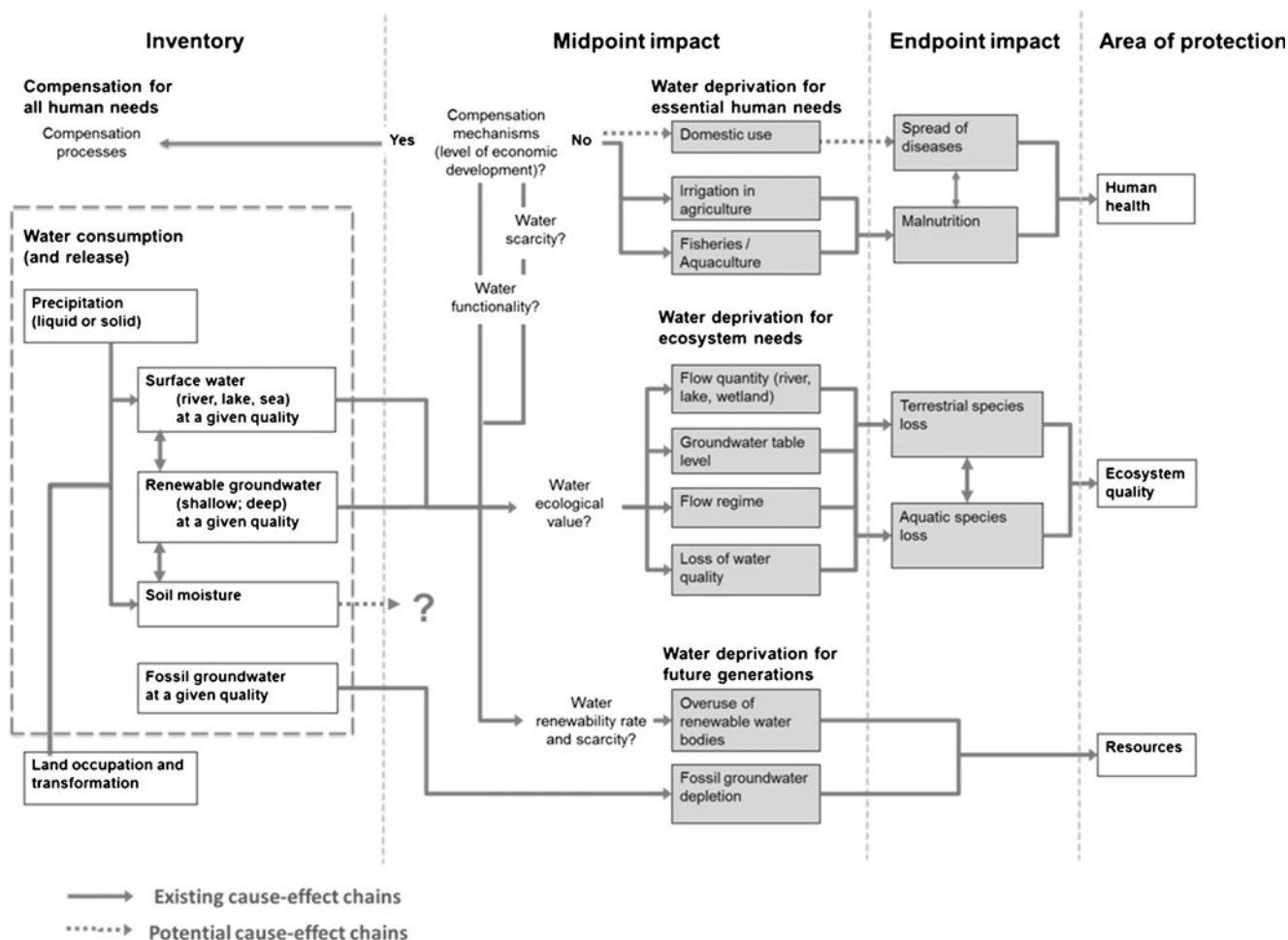


Fig. 1 Cause-effect chains leading from the inventory to the areas of protection of human health, ecosystem quality, and resources (adapted from Bayart et al. 2010)

surface water (river, lake, and sea), (2) groundwater (renewable, shallow, and deep) that is only reached through surface water and soil moisture, and (3) precipitation water stored as soil moisture (also called “green water”) (Falkenmark and Rockstrom 2006). Fossil groundwater compartment is not connected to other freshwater compartments. Freshwater is characterized by less than 1,000 ml/l of dissolved solids (USGS 2012) and encompasses all previously mentioned three types. The impact of freshwater use is related to (1) consumption of one of these water types and (2) withdrawal of one of these water types and release of surface water. Impact of degradative use is considered as withdrawal of surface or groundwater at a given quality followed by release at another quality. However, impact of direct pollutant release in freshwater and resulting cause-effect chains are excluded from the scope of this study, in which there is no value judgment regarding the inclusion of degradative use in considered methods. Related impact assessment approaches are assessed in the ILCD handbook (JRC-IES 2011). Land occupation and transformation as well as rainwater harvesting are a driver for a change in surface water and precipitation

water stored as soil moisture. The availability of the latter water type leads to debated potential impacts that are not considered in this work. However, the modification of the hydrological balance following land transformation or occupation is accounted for in the present framework as it corresponds to a modification of the amount of water that reaches the groundwater and surface water (equivalent to a consumption of the corresponding water).

The use of freshwater can generate potential impacts to humans, the ecosystems, and resources. These impacts can be related to water scarcity, water functionality, water ecological value, and water renewability rate and are influenced by the possibility to develop compensation mechanisms. Water scarcity is defined in this work being the water use approaching or exceeding the natural regeneration of water in a given area, e.g., a drainage basin. In this article, water scarcity is considered as a parameter leading to freshwater deprivation by limiting freshwater availability. Freshwater quality is defined as a set of parameters considered to characterize the chemical, physical, and biological properties of freshwater. It is related to a functionality approach,

which assesses to which users the freshwater withdrawn and released are functional (Bayart et al. 2010), and can also lead to water deprivation when water of a given quality is not available anymore for specific users. Water ecological value describes the physical relation to and dependency of ecosystems on freshwater (Bayart et al. 2010). Water renewability rate is the natural rate at which the resource is recharged. Compensation mechanisms refer to the use of backup technologies by human users deprived of “functional” freshwater to meet their needs (Boulay et al. 2011b).

2.1.1 Human health

The way human health is affected by freshwater use depends on the level of economic development and welfare (Boulay et al. 2011b; Bayart 2008). If this is sufficient, the lack of freshwater will be compensated by the development of backup technologies [such as desalination or the import of water-intensive goods as virtual water (Allan 1996)]. These compensation activities need to be assessed with a new inventory and can, in turn, lead to environmental impacts via other interventions involved in the compensation activities (e.g., climate change impacts caused by energy consumption for desalination). If the level of economic development is not sufficient to cover these costs, freshwater use will lead to water deprivation for primarily three functions which fulfill essential human needs depending on local conditions: domestic use (hygiene and ingestion), agriculture, and aquaculture/fisheries. Industrial functions of freshwater closed to human essential needs (e.g., house building and provision of pharmaceuticals) are not considered in this framework because they are more likely to consider compensation strategies rather than suffering from freshwater deprivation (Boulay et al. 2011b). Water quality degradation leads to water deprivation when it creates a loss of functionality for users who need water at a higher quality level than the released one. Users who are able to use freshwater at that or a lower quality level would not be deprived. The extent of water quality degradation depends on the amount and intensity of chemical, biological, and thermal pollution withdrawn and is related to the sanitation capacity. The withdrawn freshwater represents an adverse impact depriving users from a given amount of water at ambient water quality; the released freshwater (negative LCI flow) results in a burden reduction by making available the same amount of water for users capable to use water at that quality. Current models agree that the way human health is affected by water use depends on the level of economic development and welfare. They acknowledge that under given conditions, water use can lead to deprivation for essential human needs such as agriculture, fisheries, and domestic use and ultimately to malnutrition and spread of diseases. However, there is currently no sufficient information to determine whether freshwater use in a

low-income water-stressed region would lower water availability for domestic users or rather only affect other users, e.g., agricultural, fisheries, or industries (Boulay et al. 2011b).

Malnutrition and spread of diseases are interconnected, i.e., malnutrition could, for example, make a person more vulnerable to the spread of diseases, and reciprocally, some enteric diseases could affect the ability to absorb nutrients and thus contribute to malnutrition. Freshwater use ultimately leads to an aggregated impact on human health, generally expressed in disability-adjusted life years (Motoshita et al. 2010a, b; Boulay et al. 2011b; Pfister et al. 2009).

2.1.2 Ecosystem quality

Water use can also affect the ecosystem, for instance, by changes in the river, lake, or wetland flow quantity (e.g., due to surface water withdrawals); changes in the level of the groundwater table (e.g., due to groundwater withdrawal); changes in flow regimes (e.g., due to turbinated water use); and loss of freshwater quality. Similarly to human health, degradation corresponds to the consumption of freshwater of a higher quality (with a higher ecological value or ecological functionality) and the release of freshwater of lower quality (with a lower ecological value, thus affecting all the ecological users needing a better water quality, but not the users able to deal with a lower quality).

It should be noted that the latter cause–effect chain is related to the deprivation of freshwater of a given quality and not to the aquatic ecotoxicity, aquatic eutrophication, and aquatic acidification impact of this degradation. The midpoint impacts related to freshwater deprivation, which depend on water scarcity and water quality, eventually lead to species diversity change in aquatic and terrestrial ecosystems. The extent of these changes depends on the ecological value of water in the considered ecosystem. Ultimate impacts on ecosystem quality are commonly expressed in potentially disappeared fraction (PDF) of species on a given surface or volume during a given time (PDF per square meter per year or PDF per cubic meter per year) (van Zelm et al. 2011; Hanafiah et al. 2011).

Milà i Canals et al. (2009) suggest that changes caused by production systems on the amount of rainwater available to other users (ecosystems) through changes in the fractions of rainwater that follow infiltration, evapotranspiration, and runoff should be included as impacts on ecosystem quality. This is closely linked to the impact of land occupation and transformation on green water availability through the variation of stock of water stored as soil moisture available for plant uptake (green water).

2.1.3 Resources

Consumption of all freshwater types as well as withdrawal and release of fossil groundwater can respectively lead to overuse

of renewable water bodies or exhaustion of nonrenewable fossil groundwater. Overuse of renewable water bodies depends on the water renewability rate. These midpoint impacts affect water flows and funds and ultimately have an effect on the resources stock. This reduction of available water affects other cause–effect chains by increasing local water scarcity. Different approaches exist to characterize the impact on resources encompassing the abiotic depletion potential given in antimony equivalents (Sb-eq) (Milà i Canals et al. 2009) at the midpoint level, and the backup technology concept expressing the resource damage in megajoules (MJ) surplus energy- (Pfister et al. 2009) or exergy-based methods given in megajoules of exergy (MJex) (Boesch et al. 2007) at the endpoint level.

2.2 Definition of criteria to evaluate the existing methods

Five scientific (1–5) criteria and one potential stakeholder acceptance (6) criterion based on the ILCD Handbook (JRC-IES 2011) were adopted within this review: (1) completeness of scope; (2) environmental relevance; (3) scientific robustness and certainty; (4) documentation, transparency, and reproducibility; (5) applicability; and (6) degree of potential stakeholder acceptance and suitability for communication in business and policy contexts. They are further described in Table S5 in the [electronic supplementary material](#).

2.3 Development of sub-criteria specific to freshwater

In addition to the six criteria mentioned above, sub-criteria specific to freshwater use were added in the criteria “completeness of scope” and “environmental relevance” as described in Table 1. For the former, sub-criteria were needed to identify which areas of protection are considered by the existing methods and which mid- and endpoints are modeled. For the latter, sub-criteria were needed to evaluate the coverage of relevant freshwater-specific cause–effect chains as depicted in Fig. 1. The level of coverage was assessed without weighting the relative importance of different cause–effect chains and related parameters, but rather by exploring how far and with which method this coverage has been performed.

2.4 Description and review of existing freshwater use assessment methods

Various methods have been developed to evaluate freshwater use in LCA. Many of them were already published or in the process of being published. All methods addressing freshwater use supported by sufficient documentation to be analyzed, i.e., a draft article, a report, etc., were considered in this paper. Unpublished methods were assessed regarding the latest information available in June 2012. Figure 2 summarizes the reviewed methods and classifies them at the inventory level,

water index level, or impact assessment level, distinguishing between mid- and endpoint assessments. It identifies those specifically addressing one area of protection or more comprehensive methods that address more than one area of protection. Databases are called according to the database name and methods according to the name of the developer for academic work, e.g., Boulay (Boulay et al. 2011b) or the industry for methodology developed within a company, e.g., Veolia. A short description of assessed methods is provided in the [supporting information](#).

2.4.1 Inventory databases

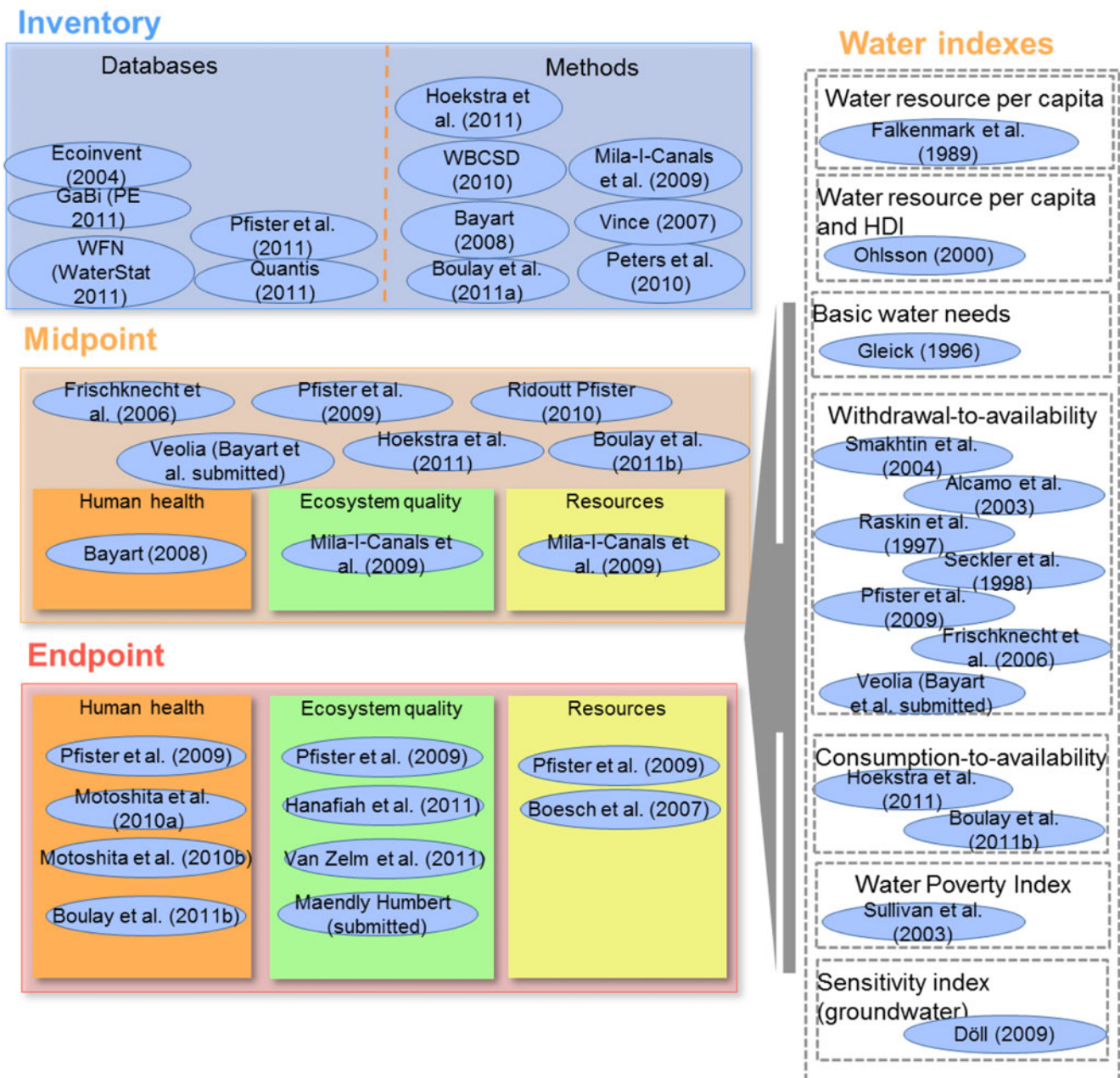
The inventory section contains both inventory databases and inventory methods. The ecoinvent database (Frischknecht et al. 2004; Ecoinvent 2007) and GaBi database (PE 2011) are widely used databases and contain elementary flows for freshwater withdrawal and turbinated water. The WFN database (Water Footprint Network 2011) assesses the inventory consumptive and degradative flows of crops and derived crop products, farm animals, and animal products; biofuels; national consumption and production; as well as trade in crop and animal and industrial products according to the WFN method (Water Footprint Network 2011). Pfister et al.’s database (Pfister et al. 2011) assesses the freshwater consumption for the production of 160 crops. An additional source of data for consumptive and evapotranspirative use can be found for five crops and three livestock products (Hanasaki et al. 2010). The Quantis water database (Quantis 2011) is a database of water uses based on ecoinvent 2.2 developed in the aim of providing industrial stakeholders with datasets required to apply all existing impact assessment methods.

2.4.2 Inventory methods

Inventory methods generally suggest concepts for a systematic classification of freshwater elementary flows according to their type (surface water, groundwater, precipitation water stored as soil moisture, whether intake water quality is considered, etc.) without providing respective data. Inventory methods also describe technical water flows such as cooling water and irrigation water. The reviewed inventory methods differ widely in their objective and level of detail. Some focus on defining water categories to allow quality to be considered (Vince 2007; Bayart 2008; Boulay et al. 2011a), and others, on providing inventory tools for organizations (Hoekstra et al. 2011; WBCSD 2010), integrating the effects of direct water use and of land occupation and transformation on water availability in a comprehensive methodology (Milà i Canals et al. 2009), or providing detailed hydrological modeling and classification of freshwater use data in specific sectors (e.g., Australian red meat sector) (Peters et al. 2010). Boulay et al. (2011a) was built on Vince’s (2007) and Bayart’s (2008) methods.

Table 1 Specific sub-criteria used to characterize inventory and mid- and endpoint modeling

Criteria	Sub-criteria	Relevant modeling aspect	
Completeness of scope	Midpoint: which impact mechanisms are covered by the impact indicators for the midpoint affecting the area of protection human health?	Water deprivation for: <ul style="list-style-type: none"> – Domestic use – Irrigation in agriculture (agricultural use) – Fisheries/aquaculture 	
	Midpoint: which impact mechanisms are covered by the impact indicator for the midpoint affecting the area of protection ecosystem quality?	<ul style="list-style-type: none"> – Changes in flow quantity (river, lake, and wetland) – Changes in groundwater table level – Change in flow regimes – Loss water quality 	
	Midpoint: which impact mechanisms are covered by the impact indicator for the midpoint affecting the area of protection resources?	<ul style="list-style-type: none"> – Overuse of renewable water bodies – Fossil groundwater exhaustion 	
	Endpoint: which impact mechanisms are covered by the endpoint indicator affecting the area of protection human health?	<ul style="list-style-type: none"> – Spread of diseases due to midpoint impact on domestic use – Malnutrition due to midpoint impact on irrigation and fisheries/aquaculture 	
	Endpoint: which impact mechanisms are covered by the endpoint indicator affecting the area of protection ecosystem quality?	<ul style="list-style-type: none"> – Terrestrial species loss – Aquatic species loss 	
	Endpoint: is the endpoint indicator affecting the area of protection resources covered?		
Environmental relevance	Water type in nature	What types of water are considered?	<ul style="list-style-type: none"> – Surface water (river, lake, and sea) – Groundwater (renewable, fossil, shallow, and deep) – Precipitation water stored as soil moisture
		Are consumption and water release considered?	
		Inventory: is intake and released water quality considered?	
	Cause–effect chain	Midpoint/endpoint cause–effect chain affecting area of protection human health: is water scarcity taken in account?	
		Midpoint/endpoint cause–effect chain affecting area of protection human health: are water functionalities of the water resource taken in account?	
		Midpoint/endpoint cause–effect chain affecting area of protection human health: are economic development level and compensation mechanisms taken in account?	
		Midpoint/endpoint cause–effect chain affecting area of protection ecosystem quality: is water ecological value taken in account?	
Midpoint/endpoint cause–effect chain affecting area of protection resources: is water scarcity taken in account?			
	Midpoint/endpoint cause–effect chain affecting area of protection resources: is water renewability rate taken in account?		



Caption: Method / index Methods or water index addressing water use

Fig. 2 Scope of and relationship between the available freshwater use inventory and impact assessment methods with classification for the three areas of protection

2.4.3 Midpoint assessment methods

Midpoint impact assessment methods give an indicator either common to all areas of protection or specific to a defined area of protection. Methods covering all area of protections giving a single index related to water scarcity include the Swiss ecological scarcity (Frischknecht et al. 2006; Pfister et al. 2009; Ridoutt and Pfister 2010b), Water Impact Index of Veolia,

Boulay et al. (2011b) methods, and Water Footprint impact indices (Hoekstra et al. 2011). Area of protection-specific midpoint indicators describe the impact pathway leading to a decrease in freshwater availability for contemporary human users (Bayart 2008), as well as changes in freshwater availability for ecosystems leading to freshwater ecosystem impacts (Milà i Canals et al. 2009) and changes in groundwater availability causing freshwater depletion (Milà i Canals et al. 2009).

Milà i Canals et al. (2009) suggest to use different types of water indices (Smakhtin et al. 2004; Falkenmark et al. 1989; Raskin et al. 1997) to assess freshwater ecosystem impacts. Falkenmark et al.'s (1989) index focuses on human use by evaluating the fraction of the total annual runoff available for human use. Raskin et al. (1997) use a water use per resource refined by Smakhtin et al. (2004) by subtracting environmental freshwater requirements from the available resources to derive a water index focused on freshwater resources available for human use.

The overall “blue-green-gray water” footprint concept of Hoekstra et al. (2011) was generally classified as an inventory metric, given that precipitation water stored as soil moisture evapotranspired by plants (“green water footprint”) and consumptive use of surface and groundwater (“blue water footprint”) represent physical metrics and are not further characterized. However, the gray water footprint can also be evaluated as a midpoint approach as gray water footprint denotes degradative freshwater use by characterizing the chemical pollution in water similar to “the critical dilution volumes approach”,¹ i.e., an equivalent amount of water needed to dilute an emission below an acceptable threshold. This method thus juxtaposes measurable inventory results of “blue” and green water footprint with a theoretical volume of “gray water” which corresponds to a characterized inventory results. Using the term gray water also creates the problem of having two competing definitions of this term circulating in the water industry² (Henriques and Louis 2011).

2.4.4 Endpoint assessment methods

Endpoint impact assessment methods provide specific indicators for potential damages on the areas of protection of human health (Boulay et al. 2011b; Motoshita et al. 2010b, a; Pfister et al. 2009), ecosystem quality (Hanafiah et al. 2011; Pfister et al. 2009; van Zelm et al. 2011), and resources (Pfister et al. 2009; Boesch et al. 2007).

Other approaches exist to estimate impact on resources that attempt to account for the energy flows put into place by natural processes to make available a given resource at a given state (Zhang et al. 2010; Rugani et al. 2011) but are not evaluated in this review because they are not specific to the characteristics of freshwater resource. Emergy is defined as the measure of both the work of nature and that of humans in generating products and services, i.e., a record of previously used-up available energy that is a property of the smaller amount of available energy in a transformed product (Odum 1996).

¹ The critical dilution volume approach characterizes each emission in terms of the volume of water required to dilute an emission to reach the legal limit set for that emission.

² In the water industry, gray water is nutrient-rich sewage from households which lacks fecal or urine contamination.

2.4.5 Water indices

Water indices are originally non-LCA-based indicators that express a measure of human and environmental water needs or of the fraction of resource available to meet these needs. Water indices can be used as characterization factors for midpoint (Raskin et al. 1997; Smakhtin et al. 2004; Falkenmark et al. 1989) and endpoint (Sullivan et al. 2003; Döll 2009) impact assessment methods when applied to freshwater consumptive or degradative use. Such indices can be considered as human use oriented (Gleick 1996; Falkenmark et al. 1989; Ohlsson 2000; Seckler et al. 1998; Sullivan et al. 2003; Döll 2009), ecosystem use oriented (Smakhtin et al. 2004), or cover all three areas of protection (Alcamo et al. 2007; Raskin et al. 1997; Pfister et al. 2009; Frischknecht et al. 2006; Hoekstra et al. 2011; Boulay et al. 2011b). In this work, the terminology “water scarcity index” is related solely to withdrawal-to-availability ratio (Smakhtin et al. 2004; Alcamo et al. 2007; Raskin et al. 1997; Seckler et al. 1998; Pfister et al. 2009; Frischknecht et al. 2006; Bayart et al. submitted) or consumption-to-availability ratio (Boulay et al. 2011b; Hoekstra et al. 2011). Water scarcity indices can be based solely on a measure of water scarcity or include, additionally, a measure of water quality (Boulay et al. 2011b). The details of the implementation of water indices in a LCA context, i.e., the water type to be considered in the inventory phase, needs to be specified in order to make water indices applicable in a method.

2.4.6 Uncertainty

Uncertainties are generally large in life cycle impact assessment, especially on the endpoint level, and are yet generally not quantified in most of methods. Only a few authors, i.e., Pfister and Hellweg (2011), reported uncertainties for human health and WSI indicators on watershed and country level.

3 Results and discussion on method evaluation and cross comparison

The methods were evaluated and compared according to the selected criteria and sub-criteria displayed in Table 1 and S5 at the inventory, midpoint, and endpoint levels, and key differences were identified. Tables S1, S2, and S3 in the [electronic supplementary material](#) provide a summary of the review for each method.

3.1 Inventory databases

While the ecoinvent (Frischknecht et al. 2004; Ecoinvent 2007), GaBi (PE 2011), and Quantis (Quantis 2011) databases give the opportunity to distinguish freshwater input as water withdrawal according to its natural source [surface water (river and lake) or groundwater (renewable and fossil)], in

the WFN database (Water Footprint Network 2011) and Pfister et al.'s datasets (Pfister et al. 2011), water input is restricted to consumption of precipitation water stored as soil moisture evapotranspired by plants (so-called green water footprint) as well as consumption of surface and groundwater combined (so-called blue water footprint). All datasets consider water outputs and global water balances in a different manner. The ecoinvent datasets in their current version 2.2 do not allow the determination of water balances because water releases are not reported, water consumption being thus an unknown part of the withdrawal. In contrast, GaBi and Quantis databases contain water inputs and outputs for all fore- and background processes. The WFN database (Water Footprint Network 2011) considers volumetric estimations of water consumption through blue and “green” water footprint, while degradative use is expressed through the gray water concept, where pollutant persistence, inter-compartment transfer, and bioaccumulation properties are only implicitly included in water quality standard definitions which exist for a reduced set of substances. Only the Quantis water database considers water evaporated from reservoirs. The WFN and GaBi databases, the Quantis water database, and Pfister et al.'s datasets provide regionalized data per country which were appropriate in regards to the product (global commodities or region-specific products).

3.2 Inventory methods

The water flow classification of Boulay et al. (2011a) and the Global Water Tool of the WBCSD (2010) distinguish water according to its origin (e.g., surface and groundwater) and account for water balances by using input–output inventories. Milà i Canals et al.'s (2009), Peters et al.'s (2010), and WFN (2011) methods account only for consumptive water use of soil moisture lost by evapotranspiration (green water) as well as evaporated surface and groundwater flows. Boulay et al.'s method, which is an upgraded version of Vince's (2007) and Bayart's (2008) methods, is more comprehensive, as it enables to classify 11 input and output water inventory flows by using corresponding water quality classes based on 137 parameters.

3.3 Midpoint assessment methods

Water indices used in midpoint methods are based on a withdrawal-to-availability ratio (Pfister et al. 2009; Ridoutt and Pfister 2010b; Frischknecht et al. 2006; Milà i Canals et al. 2009) or a consumption-to-availability ratio (Boulay et al. 2011b; Hoekstra et al. 2011). They are used as a characterization factor for freshwater use in life cycle impact assessment to assess the impact of water consumption (Pfister et al. 2009; Ridoutt and Pfister 2010b; Frischknecht et al. 2006; Boulay et al. 2011b; Hoekstra et al. 2011; Milà i Canals et al. 2009) and water degradation (Hoekstra et al. 2011; Boulay et al. 2011b;

Ridoutt and Pfister 2010b). Ridoutt and Pfister's (2010b) index is an extended version of Pfister et al.'s approach (2009), given that degradative water use (gray water) is included additionally to consumptive use (“blue water consumption”). The Water Footprint Network's impact indices (green, blue, and gray water footprint impact indices) (Hoekstra et al. 2011) follow the same concept by applying blue, gray, and green water scarcity indices to corresponding water categories. The Water Impact Index and Boulay et al.'s (2011b) index both include water quality as a parameter, additionally, to water scarcity considering that water quality parameters could restrict its use by humans and the natural environment as defined in Fig. 1. Storage capacity has been considered in Pfister et al. (2009) as it is strongly related to water deprivation (deprivation occurs only if storage capacity is insufficient or if much of the stored water is evaporated).

Most of the methods provide characterization factors differentiated by country (Frischknecht et al. 2006; Bayart 2008; Pfister et al. 2009; Milà i Canals et al. 2009; Ridoutt and Pfister 2010b; Boulay et al. 2011b), watershed (Frischknecht et al. 2006; Water Footprint Network 2011; Boulay et al. 2011b; Pfister et al. 2009), or grid cell (Pfister et al. 2009; Ridoutt and Pfister 2010b). The Water Footprint Network blue water footprint impact indices provide characterization factors with monthly temporal differentiation (Hoekstra et al. 2011) and thus offer more temporal precision for impact evaluation. However, storage of water is not included. Milà i Canals et al. (2009) methods on freshwater depletion do not provide regionalized characterization factors.

3.4 Endpoint assessment methods

3.4.1 Human health

The impact pathways covered by current methods regarding human health include the lack of freshwater for hygiene and ingestion resulting in the spread of communicable diseases (Motoshita et al. 2010b; Boulay et al. 2011b), water deprivation for irrigation causing in malnutrition (Pfister et al. 2011; Motoshita et al. 2010a; Boulay et al. 2011b), and water deprivation for freshwater aquaculture and fisheries resulting in loss of productivity and food supply (Boulay et al. 2011b). Indirect impact of freshwater use, i.e., impact on human health and conflict creation, is not covered by existing methods. The cause–effect chain modeling is based on hydrological and socioeconomical data (Pfister et al. 2009; Boulay et al. 2011b; Motoshita et al. 2010b, a). Some of them consider the water scarcity index used at the midpoint level (Pfister et al. 2009; Boulay et al. 2011b). The level of economic development is considered in studied methods through parameters such as Human Development Index (Pfister et al. 2009), house connection to water supply (Motoshita et al. 2010b), or adaptation capacity based on

gross national income (Boulay et al. 2011b). All methods consider the reduction of human health impacts in case the level of economic development is sufficient to cover compensation mechanism costs, but none of them include the impact of the development and functioning of compensation mechanisms. Not expanding the system boundary is a common approach in attributional LCA. Some of the cause–effect chain relationships have been calculated based on empirical data, e.g., malnutrition rate and human development index (Pfister et al. 2009), water scarcity, and accessibility to safe water (Motoshita et al. 2010b). Other cause–effect chains rely on the multiplication of key parameters (Boulay et al. 2011b). Both approaches are relevant but need to be further characterized by a measure of uncertainty to assess the deviation of potential impacts estimation. Endpoint indicators are generally regionalized on a country (Pfister et al. 2009; Motoshita et al. 2010b; Boulay et al. 2011b) or watershed level (Pfister et al. 2009; Boulay et al. 2011b).

3.4.2 Ecosystem quality

Methods addressing ecosystem quality cover different parts of the cause–effect chains relevant to ecosystem services and biodiversity. The cause–effect chains that current methods cover regarding damages to ecosystem quality are the decrease of terrestrial biodiversity due to freshwater consumption (Pfister et al. 2009), decrease of aquatic biodiversity due to turbined water use, disappearance of terrestrial plant species due to groundwater withdrawal and related lowering of the water table (van Zelm et al. 2011), and the effects of freshwater consumption on freshwater fish species (Hanafiah et al. 2011). These endpoint methods do not use water scarcity indices as elements of the modeling equations. Rather, they are applied to different water types and uses and should be used complementarily. Most methods consider the ecological value of freshwater resources through an empirical observation of decreased biodiversity or of other proxy data such as net primary production (Pfister et al. 2009; van Zelm et al. 2011) and from a mechanistic perspective, e.g., by relating fish species richness to river discharge (Hanafiah et al. 2011).

Some cause–effect chains, e.g., the impact due to water deprivation related to water quality degradation on aquatic ecosystems, still need to be covered by additional methods. Endpoint methods addressing ecosystem quality have different levels of spatial differentiation: no differentiation, generic or for a specific region (van Zelm et al. 2011), archetype (e.g., alpine and non-alpine dams) country (Pfister et al. 2009), or watershed (Pfister et al. 2009; Hanafiah et al. 2011). This variability of the differentiation level reflects the diversity of the parameters considered in the cause–effect chain.

3.4.3 Resources

Methods addressing the area of protection resources quantify the impact on future freshwater availability through a backup technology approach to evaluate the impact of freshwater consumption above their renewability rate (Pfister et al. 2009) or through the exergy content of the freshwater resource (Boesch et al. 2007). In contrast to the Pfister et al.'s method (2009), Boesch et al.'s (2007) method is not specific to water resources and does not consider water scarcity.

None of the evaluated endpoint methods cover the cause–effect chain comprehensively; the pathway addressing impact due to fossil groundwater depletion is poorly known and is not covered by available methods. Furthermore, estimation of impact of consumption over the renewability rates lacks differentiation between different water types, and change in green water availability is not covered. Pfister et al.'s (2009) method is a spatially differentiated method on a watershed and a country level, whereas Boesch et al.'s (2007) method is not differentiated.

4 Recommendations

4.1 Model components to build a scientific consensus for method developers

The previously described findings can guide future consideration of freshwater use in LCA.

4.1.1 Inventory databases

From a business and industry perspective, data availability on freshwater use as well as harmonized reporting formats are limiting factors for establishing meaningful water footprints of products, processes, and organizations (Koehler 2008). A balanced approach between LCIA methods and business data requirements is therefore needed to make characterization methods broadly applicable and meaningful. In order to link up with emerging LCI and LCIA methods, inventory databases should preserve the maximum freedom to provide necessary flows for application of different impact methods. The following recommendations for inventory database developments were drawn based on existing LCI and LCIA methods and are evaluated as necessary:

1. Differentiate consumptive freshwater use from withdrawal (abstraction) through consistent water balances for fore- and background processes and do not mix physical flows with assessment units such as cubic meter equivalents of polluted water.
2. Distinguish between different water types based on origin (surface freshwater, including river, lake, and

sea; groundwater, including renewable, shallow, and deep; and precipitation freshwater stored as soil moisture and fossil groundwater) and freshwater quality (and thus functionality). This can be done by applying the systematic classification proposed by Boulay and colleagues (2011a) according to quality data that could be collected, e.g., data on general parameters (which include microbial parameters) and inorganic and organic compounds.

3. Include freshwater evaporation from water reservoirs as consumptive use, as it makes freshwater locally/regionally not available anymore.

The following additional optional guidelines could be integrated:

1. For the assessment of groundwater withdrawals and associated impacts, differentiate shallow (<2.3 m) and deep water tables (van Zelm et al. 2011) or estimate regional average fractions of areas of each type.
2. Differentiate withdrawal of fossil groundwater from renewable groundwater based on regionally available resources as far as possible.

4.1.2 Inventory methods

General recommendations for inventory methods are the following:

1. Include only measurable freshwater types, e.g., surface water and groundwater, or a method to estimate those flows shall be provided, e.g., for water stored as soil moisture evapotranspired by plants, so called green water, which can be estimated with a crop model suitable, based on input data on climate, soil, and crop characteristics (Hanasaki et al. 2010; Hoekstra et al. 2011).
2. Use water quality parameters to characterize freshwater flows that are available in existing databases.

4.1.3 Impact assessment methods

In order to ease their applicability, LCIA shall, in general, show robust examples linking the inventory of freshwater types with all needed calculation steps to apply characterization factors and aggregate results for obtaining related mid- or endpoint indicators.

4.1.4 Midpoint assessment methods

The water consumption or withdrawal to availability ratio has been recognized as a representative proxy for scarcity, in comparison to other indices, e.g., water use per capita, which reflects rather a socioeconomic situation. Midpoint methods

addressing water scarcity shall (1) include water storage capacity in the modeling of total water availability within a geographical unit, (2) be quantitatively compared to evaluate the trade-off between easiness of application and cause–effect chain coverage and related uncertainty between indicators based solely on water scarcity (Pfister et al. 2009; Ridoutt and Pfister 2010b; Frischknecht et al. 2006; Milà i Canals et al. 2009) and more comprehensive midpoint indicators (Boulay et al. 2011b), (3) provide further empirical evidence of the link between water scarcity, water deprivation, and impact on different areas of protection to evaluate the relevance of mid- versus endpoint indicators. In an LCA perspective, water scarcity indicator does not refer to any potential impact. This does not necessarily mean that an endpoint is ultimately affected. Water stress index is, for example, involved in Pfister et al.'s and Boulay et al.'s endpoint models for human health, but human health is not affected if the economic development level is sufficient. Clear evidence of the link between water scarcity, water deprivation, and impact on different areas of protection would be needed to evaluate the relevance of mid- versus endpoint indicators.

4.1.5 Endpoint assessment methods

Next steps towards a consistent framework for application of endpoint methods are as follows.

1. For the area of protection *human health*:
 - a. Provide a quantitative comparison of existing methods as well as an evaluation against empirical figures.
 - b. Assess the relevance and uncertainty of modeling indirect impacts related to water deprivation, e.g., human health impact due to conflict creation, and population displacement.
 - c. Develop new approaches for modeling of compensation mechanisms to prevent water loss in functionality throughout impact categories, knowing that technical means can also be used to cope with other impacts such as climate change.
2. For the area of protection *ecosystem quality*:
 - a. Identify extensively missing cause–effect chain.
 - b. Provide global coverage and appropriate spatial resolution (e.g., watershed scale) for methods developed for a single country (van Zelm et al. 2011) or with partial basin coverage (Hanafiah et al. 2011).
3. For the area of protection *resources*:
 - a. Cover the cause–effect chain leading to impact of fossil groundwater exhaustion, as well as include it in the inventory.

- b. Distinguish impact related to different freshwater types consumption, given that they have different renewability rates and functionalities.
- c. Quantifying the link between green water use and resources. Although Heuvelmans et al. (2005) developed a method to quantify impact of land use concerned with changes in hydrological response of the land, no characterization factors yet exist to quantify this relationship.

For all mid- and endpoint methods, uncertainties of input data as well as model uncertainty still need to be evaluated and documented. Mid- and endpoint methods covering human health and ecosystem quality impact shall provide characterization factors with monthly differentiation to reflect variability related to meteorological conditions and associated ecosystem changes.

4.2 Application recommendations for practitioners given current state of the art

The evaluation of freshwater use is possible by assembling methods in a comprehensive methodology to characterize each use adequately. Current state of the art can already provide a preliminary understanding of water uses and associated impacts, especially on human health and ecosystem quality.

In this respect, a detailed inventory, including freshwater withdrawal and release, water consumption, and turbinated water, constitutes a first step towards understanding the various flows related to the system. Inventory results can be used as an indicator as such (Hoekstra 2011; WBCSD 2010), but the interrelation between inventory results and impact linked to freshwater use is not yet proven and can be, in some cases, misleading (Ridoutt et al. 2010; Ridoutt 2011). Clarity of LCI scope demands clear communication regarding whether an attributional or consequential LCI approach has been taken.

For midpoint level assessment, it is not yet possible to draw conclusions on method preference, given that case studies to test the significance of each method are under development. It is recommended to use the existing midpoint methods most relevant for the study under elaboration to ensure an extensive sensitivity analysis on the methodological choice, keeping in mind their cause–effect chain overlaps. If possible, the information given by scarcity indices should be interpreted in parallel with damage-oriented impact assessment indicators to provide a comprehensive picture of impacts related to freshwater use.

For endpoint level assessment, it is recommended to combine indicators of all cause–effect chains, i.e., malnutrition or infectious diseases related to water deprivation of a defined quality class for agriculture, fisheries, and domestic use for human health. For ecosystem quality, the scopes of

methods developed could so far be considered as complementary. All ecosystem quality indicators could therefore be used simultaneously and summed up into a single metric. However, results should be interpreted with caution as not all the indicators are addressing the same endpoints. The resource area of protection is considered not being sufficiently developed to provide significant results.

This assessment needs to be completed by emission to all compartments ultimately affecting water (e.g., aquatic acidification, eutrophication, human toxicity, ecotoxicity, as well as heat release to water) to provide a complete picture of water-related impacts.

5 Conclusions

This is the first state-of-the-art assessment of freshwater use-related methods. This review assesses relevant tools to make an assessment from a product or site perspective, extending the analysis beyond the water flow inventory and encompassing impact from indirect water use in the system limits. Although some cause–effect chains still need to be covered, spatial differentiation refined, and uncertainty assessed, the set of methods presented can already help to grasp water-related challenges and risks which humans face and serve as a first base for strategic decisions. Water assessment is a fast progressing field, and this review will need to be regularly updated to include new developments.

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